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Cite this article: Donelan SC, Trussell GC.

2018 Parental and embryonic experiences with predation risk affect prey offspring behaviour and performance. *Proc. R. Soc. B* 20180034.

<http://dx.doi.org/10.1098/rspb.2018.0034>

Received: 5 January 2018

Accepted: 16 February 2018

Subject Category:

Ecology

Subject Areas:

ecology

Keywords:

parental effects, embryonic effects, developmental plasticity, transgenerational plasticity, life history, *Nucella lapillus*

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Electronic supplementary material is available online at <https://dx.doi.org/10.6084/m9.figshare.c.4015747>.

Parental and embryonic experiences with predation risk affect prey offspring behaviour and performance

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Because phenotypic plasticity can operate both within and between generations, phenotypic outcomes are often shaped by a complex history of environmental signals. For example, parental and embryonic experiences with predation risk can both independently and interactively influence prey offspring traits early in their life. Parental and embryonic risk experiences can also independently shape offspring phenotypes throughout an offspring's ontogeny, but the persistence of their interactive effects throughout offspring ontogeny is unknown. We examined the effects of parental and embryonic experiences with predation risk on the response of 1-year-old prey (the carnivorous snail, *Nucella lapillus*) offspring to current predation risk. We found that parental and embryonic risk experiences had largely independent effects on offspring performance and that these effects were context dependent. Parental experience with risk had strong impacts on multiple offspring traits in the presence of current risk that generally improved offspring performance under risk, but embryonic risk experience had relatively weaker effects and only operated in the absence of current risk to reduce offspring growth. These results illustrate that past environmental experiences can dynamically shape organism phenotypes across ontogeny and that attention to these effects is key to a better understanding of predator/prey dynamics in natural systems.

1. Introduction

Organisms frequently respond to changes in their environment via phenotypic plasticity, and such modifications have clear implications for individual fitness [1–3] and community and ecosystem dynamics [4,5]. Phenotypic expression can also be influenced by an individual's previous or historical experiences that persist to affect future performance (e.g. carryover effects [6]). One particularly influential period is embryonic development [7]. Both theoretical and empirical work indicates that the environmental conditions experienced as an embryo can affect an organism's phenotype at emergence [8,9], but also have outsized impacts on its lifetime fitness trajectory [10–12]. Moreover, embryonic experience can persist through major life-history shifts such as metamorphosis [13] and because of its influence on adult traits, have lasting impacts on multiple generations [14].

Organisms can also exhibit phenotypic plasticity in response to their parents' environmental experience via parental effects (i.e. transgenerational phenotypic plasticity [15–17]). Parental effects may be adaptive for offspring (and therefore parents) if they appropriately anticipate future conditions and hence mitigate uncertainty in heterogeneous environments [18,19]. Adaptive parental effects are, therefore, likely to be more prevalent in systems where a parent's environment is a strong predictor of its offspring's environment and when environmental conditions vary over ecologically relevant time scales [18,20]. Such parental effects operate in a variety of systems and in response to a diverse array of environmental cues, even in species without parental care (e.g. predation risk [21], environmental quality [22] and climate change [23]). Parental effects can

also have remarkable longevity throughout the lifetime of offspring [24,25] and extend beyond one generation (grandparental effects [26]). Hence, the influence of parental experience on offspring is probably a pervasive feature of natural systems that has important effects on phenotypic expression.

While parental and embryonic effects each have independent and lasting impacts on individual fitness, they may also interact to influence individual performance. Reinforcement between parental and embryonic environments, for example, may more reliably indicate the conditions that an individual is likely to encounter than either signal alone. Parental and embryonic effects are more likely to interact in systems where offspring have low dispersal potential and thus develop in conditions similar to their parents or where species have short life cycles relative to the rate of environmental change [12]. Despite the potential for parental and embryonic experiences to interact, our understanding of how and when this interaction operates in offspring and whether it changes across offspring ontogeny is quite limited. Theory predicts that more recent environmental signals should have a stronger influence on organismal traits because the reliability of information obtained from a changing environment declines over time [27]. Hence, closer temporal proximity between the parental and embryonic environments and those experienced immediately after emergence may increase the likelihood that parental and embryonic effects have greater impacts early in offspring ontogeny. There is evidence for ontogenetic shifts in the independent influence of parental (e.g. [28,29]) and embryonic (e.g. [30]) effects. For example, bryozoans whose mothers experienced high loads of copper pollutant performed better as larvae when exposed to copper themselves, but these effects changed over time such that maternal effects diminished offspring performance at later life stages, particularly in stressful conditions [31]. Thus, the nature and strength of any interactive effects between parental and embryonic experience may also change over an individual's lifetime.

Predation risk, where predators scare rather than consume their prey, can be a strong driver of prey behaviour and performance in many systems. In the presence of predation risk, prey often seek refuge in safe habitats to reduce their risk of being consumed [32,33], but by doing so their foraging and performance can suffer [34,35]. These effects can cascade throughout the community and ecosystem with important consequences for population dynamics [36], resource abundance [37] and nutrient cycling [38]. Furthermore, both parental and early life experiences with predation risk can independently affect offspring traits at emergence [9,39] and later in life [40,41]. Donelan & Trussell [42] also found that parental experience with predation risk impacted the size of prey offspring at emergence in a rocky intertidal snail, but only if the offspring were also exposed to risk as embryos. Whether this interaction between parental and embryonic experience with predation risk persists throughout offspring ontogeny, however, remains unknown.

The snail *Nucella lapillus*, an important intermediate consumer on rocky shores, increases its use of refuge habitats [43], reduces its foraging behaviour [44], produces less tissue [45] and grows less efficiently [46] in the presence of predation risk from the green crab (*Carcinus maenas*). Parental experience with green crab predation risk, however, can reduce these fitness consequences in adult *Nucella* offspring [47] and, when combined with embryonic risk exposure, reduce the physiological costs of predation risk in offspring as embryos, thereby

allowing them to achieve a larger size at emergence [42]. To explore whether parental and embryonic exposures to predation risk interact to influence the response (behaviour and performance) of offspring to predation risk later in life history, we conducted a laboratory experiment where we manipulated parental experience with predation risk, embryonic experience with risk and exposure to current risk in 1-year-old, subadult *Nucella*. Our results suggest that the strength and nature of parental and embryonic effects can change during offspring ontogeny and that organisms can dynamically integrate past environmental experiences into current phenotypic outcomes. Parental effects, however, appear to have lasting and substantial impacts on prey performance and probably play an important role in predator/prey dynamics in natural systems.

2. Material and methods

We examined the effects of parental experience (presence/absence) and embryonic experience (presence/absence) with predation risk from the green crab *C. maenas* on the response of 1-year-old *N. lapillus* (a carnivorous snail, hereafter *Nucella*) offspring to current green crab predation risk (presence/absence). Offspring were born and raised in the flow-through seawater facilities at the Marine Science Center in Nahant, MA, USA. Parent *Nucella* (males and females > 20 mm shell length [48]) were collected from an exposed rocky intertidal shore in Nahant, MA, USA, in early February 2015, returned to the Marine Science Center, and held separately by sex until late spring to allow females to expel any stored sperm (*Nucella* can store sperm for up to three months [49]) prior to experimental mating. Importantly, we collected snails of a similar age class (i.e. size) and from a relatively small area (approx. 20 m²) at the same site. *Nucella* are direct developers and are not highly mobile [50] so individuals in a small area have probably experienced similar environmental conditions throughout their lives.

In mid-May, male and female *Nucella* were randomly paired to create 50 mating pairs. Our experimental design independently manipulated parental and embryonic risk environments, so it was necessary to prevent females from depositing egg capsules in the presence of risk. We, therefore, created a week-long mating cycle consisting of two stages: (i) the risk manipulation stage, where parent snails were placed in the presence or absence of predation risk for 3 days, but kept separately and thus not allowed to mate, followed by (ii) the mating stage, where the male and female in a given pair were placed together in the same chamber in the absence of risk for 4 days to mate and deposit egg capsules.

In the risk manipulation stage, each male and female in a given pair was placed in its own perforated jar (8 × 10 cm, dia. × h) with six blue mussels (*Mytilus edulis*, 13.8 ± 1.4, mean shell length ± s.d.) for food. Jars containing the male and female in a given pair were placed together in a larger plastic bucket (24 × 24 cm, dia. × h) that was independently supplied with flowing seawater and also contained a perforated 'risk manipulation' chamber (11.5 × 10 cm, dia. × h). This risk manipulation chamber housed either one male green crab (73.7 ± 2.2 mm, mean carapace width ± s.d.) with two *Nucella* for food (risk present) or two *Nucella* alone (risk absent). Importantly, *Nucella* are highly sensitive to the presence of *Carcinus* even without the presence of food snails (e.g. [46]) and do not respond to the presence of other large intertidal crab species [51]. Food mussels and food snails were replenished each week. After 3 days in the risk manipulation stage, the male and female in each parent pair were moved to a separate mating stage bucket (24 × 24 cm, dia. × h) where they were placed together in the same perforated mating chamber (11.5 × 10 cm, dia. × h) that received an independent supply of flowing water. Parents remained in the mating stage for 4 days

to mate and deposit egg capsules, hence all egg capsule deposition occurred in the absence of risk. After 4 days, each pair was re-separated and placed in their original risk manipulation bucket, as before. This cycle continued for 12 weeks. Because there were 50 parent pairs, there were 50 independent buckets in the risk manipulation stage and 50 independent buckets in the mating stage.

Mating chambers were inspected each week for newly deposited egg capsules. If we found newly deposited egg capsules, this new clutch was removed and divided approximately in half; each half was then placed into its own mesh-lined tea infuser (Upton Tea Imports). One tea infuser from each clutch was placed into its own bucket (17×14 cm, dia. \times h) that also contained a perforated chamber ($10 \times 10 \times 7$ cm, $l \times w \times h$) that housed one adult male green crab (presence of embryonic risk). The other tea infuser from that clutch was placed into its own bucket which contained an empty perforated container (absence of embryonic risk). Each new clutch was similarly divided so that each embryonic risk manipulation bucket contained only one tea infuser; there were a total of 75 independent embryonic risk manipulation buckets ($n = 39$ for presence of embryonic risk, $n = 36$ for absence of embryonic risk). All egg capsules remained in the presence or absence of embryonic risk for one week. After one week, egg capsules were moved to a new tea infuser that was placed in its own plastic jar (8×10 cm, dia. \times h) that received risk-free, flowing seawater.

Six weeks after egg capsule deposition, we began to inspect tea infusers every 2–3 days for the emergence of *Nucella* offspring. Newly emerged offspring (1.2 ± 0.1 mm, mean shell length \pm s.d.) were given approximately 200 juvenile blue mussels (1.1 ± 0.2 mm, mean shell length \pm s.d.) for food immediately after emergence and each week thereafter until they were large enough to switch to larger mussels (4.6 ± 1.9 mm, mean shell length \pm s.d.). Before the onset of winter, offspring from each tea infuser were transferred to perforated plastic jars (8×10 cm, dia. \times h) that were placed in a larger plastic bucket that, owing to logistical constraints, also contained five other jars that held offspring from the same parental experience \times embryonic experience treatment combination. Each bucket received its own supply of seawater and ‘winter bucket’ had no effect on initial shell or tissue mass of offspring ($p > 0.1$). *Nucella* offspring were fed an ad libitum supply of blue mussels and held in these risk-free conditions until the following summer. The offspring used in this experiment are siblings of those used in our earlier work [42].

The following July, approximately 1 year after their emergence from egg capsules, we exposed *Nucella* offspring to the presence and absence of current predation risk in a laboratory mesocosm experiment at the Marine Science Center. *Nucella* offspring were 14–19 mm in shell length and, therefore, were considered subadults [45] that were approaching sexual maturity [48]. Mesocosms ($27 \times 15 \times 5$ cm, $l \times w \times h$) consisted of two chambers separated by a perforated wall: an upstream chamber with a perforated roof for the manipulation of predation risk and a downstream chamber that housed experimental *Nucella* offspring and their food. We manipulated exposure to current predation risk by placing one male green crab (75.2 ± 3.7 mm, mean carapace width \pm s.d.) with two *Nucella* food snails (risk present) or two *Nucella* food snails alone (risk absent) in the upstream chamber. The downstream chamber held four experimental *Nucella* offspring (16.3 ± 1.6 mm, mean shell length \pm s.d.) from the same treatment combination (parental experience \times embryonic experience) and 60 blue mussels (13.0 ± 2.0 mm, mean shell length \pm s.d.) for food. Mussels were placed on top of a granite tile ($15 \times 15 \times 1$ cm, $l \times w \times h$) that was elevated on 1 cm PVC spacers to create a narrow space under the tile to provide a refuge for *Nucella* offspring [44]. Each mesocosm was placed in a larger plastic box ($33 \times 19 \times 12$ cm, $l \times w \times h$) that received an independent supply of flowing seawater. There were eight replicates for each treatment combination ($n = 64$), and the experiment ran for 25 days. Offspring

from 23 parent pairs (10 pairs in the presence of parental risk, 13 in the absence of parental risk) were distributed evenly among the mesocosms as appropriate, and 6–12 (mean = 8.4) families were represented in each treatment combination.

Every 3–4 days, we monitored *Nucella* offspring refuge use in each mesocosm by recording the location of each snail. Offspring were considered in refuge if they were found underneath the tile, while all other locations were considered risky habitat. We calculated the proportion of *Nucella* offspring in refuge in each mesocosm by counting the number of snails found in refuge during a given observation and dividing by the total number of snails in that mesocosm ($n = 4$). We made a total of seven behavioural observations. We measured *Nucella* offspring tissue growth (final – initial tissue, g) by marking each snail with a numbered bee tag and weighing them using a non-destructive buoyant weighing technique [52] at the beginning and end of the experiment. We converted tissue growth to its energetic equivalent (Joules, J) using empirically derived equations (electronic supplementary material, appendix S1) that convert measured tissue growth into dry tissue mass (milligrams, mg [53]) and dry tissue mass into its energetic equivalent (J, [50]). We determined *Nucella* foraging activity by counting the number of mussels consumed (indicated by a drill hole on remaining shell) and used the maximum shell length (millimetres, mm) of each consumed mussel to calculate its dry tissue weight (mg [54]) and tissue energetic value (19.5 J mg^{-1} [55]). *Per capita* *Nucella* offspring foraging activity was then determined by dividing the total amount of energy (Joules) consumed by all offspring in a given replicate by the number of offspring in that replicate ($n = 4$).

We calculated *Nucella* offspring growth efficiency by dividing individual tissue growth (Joules) by the average *per capita* foraging activity (Joules) in that replicate. Growth efficiency is a measure of an individual’s ability to convert ingested energy into body mass and while it integrates changes in growth and foraging, it can also be directly reduced by predation risk [46]. We focused on tissue growth because tissue is more energetically expensive to produce than shell in *Nucella* [56] and is, therefore, a better indicator of an individual’s total energetic requirements.

We analysed *Nucella* offspring refuge use and foraging activity using separate type III ANOVAs that considered parental experience with predation risk, embryonic experience with risk and current exposure to risk as fixed effects. Refuge use and foraging activity analyses were done on replicate averages ($n = 64$); for refuge use, we used the average proportion of offspring in refuge during the seven observations. Because we cannot account for individual offspring foraging rates, we calculated the *per capita* foraging rate for each replicate and applied it to all offspring in that replicate.

We analysed individual *Nucella* offspring tissue growth (Joules) and growth efficiency using separate split-plot type III ANOVAs that considered parental experience with predation risk, embryonic experience with risk and current risk exposure as fixed effects and included parent pair (i.e. family) nested within parental experience with risk as a random effect to account for potential differences in the response of parent pairs to predation risk. Because there were multiple *Nucella* in each replicate, replicate was considered a random effect nested within the parental, embryonic and current risk treatments. Finally, embryonic bucket ID was included as a random block effect.

We conducted the analyses in JMP 11 using REML variance estimates and explored any significant interactions using least-square (LS) contrasts to compare group means. Two replicates (–parental risk/–embryonic risk/+current risk and +parental risk/+embryonic risk/+current risk) were excluded because half of the offspring died mid-way through the experiment. We explored the significance of the random effects using likelihood ratio tests ([57]; see the electronic supplementary material). Data

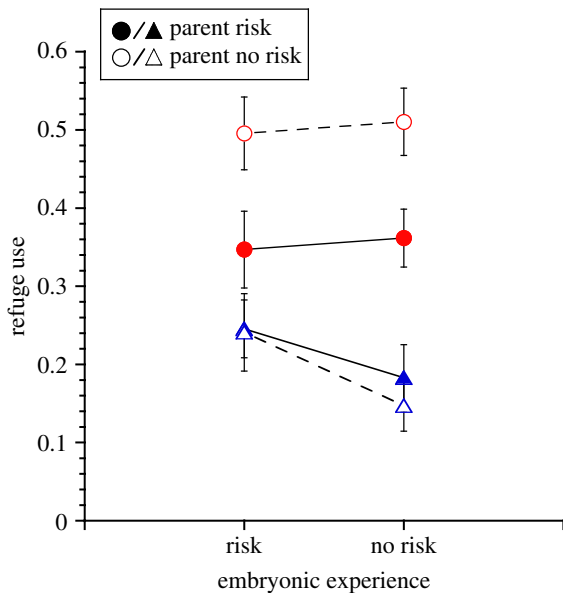


Figure 1. Mean (\pm s.e.) refuge use of offspring *Nucella lapillus* in the presence (circles) and absence (triangles) of current predation risk from the green crab *Carcinus maenas*. Offspring experienced the presence and absence of green crab predation risk as embryos and were produced by parents that experienced the presence (filled symbols) and absence (open symbols) of green crab predation risk. The three-way interaction is not significant here (parental \times embryonic \times current: $p = 0.8$), but is shown for ease of comparison to figure 2. $n = 8$ for all treatment combinations except: – parental risk/– embryonic risk/+ current risk and + parental risk/+ embryonic risk/+ current risk. (Online version in colour.)

are available in the Dryad Digital Repository [58], and ANOVA and likelihood ratio test results are provided in the electronic supplementary material.

3. Results

In the presence of current risk, *Nucella* offspring used refuge more often (current risk: $F_{1,54} = 55.7$, $p < 0.0001$; figure 1), but the offspring of risk-experienced parents used refuges 29% less than offspring of risk-naive parents (parental experience \times current risk: $F_{1,54} = 7.9$, $p = 0.007$, LS contrast: $p = 0.001$; figure 1). Refuge use was not affected by embryonic risk experience ($F_{1,54} = 1.1$, $p = 0.3$).

Despite differences in refuge use, there were no effects of parental or embryonic experiences with risk on offspring foraging activity (all $p > 0.14$), but offspring did forage less in the presence of current risk (current risk: $F_{1,54} = 466.7$, $p < 0.0001$; electronic supplementary material, figure S1). By contrast, parental, embryonic and current risk experiences interactively influenced offspring tissue growth (parental experience \times embryonic experience \times current risk: $F_{1,53.3} = 4.3$, $p = 0.043$; figure 2). In the presence of current risk, offspring of risk-experienced parents grew 128% more tissue than offspring of risk-naive parents (LS contrast: $p = 0.05$; figure 2), but there was no effect of parental risk experience in the absence of current risk (LS contrast, $p = 0.8$). In the absence of current risk, embryonic risk experience influenced tissue growth for offspring of risk-naive parents (LS contrast: $p = 0.01$) but not for offspring of risk-experienced parents (LS contrast $p = 0.6$). Offspring of risk-naive parents that did not experience risk as embryos grew 18% more tissue than offspring that did experience risk as embryos. There was no effect of

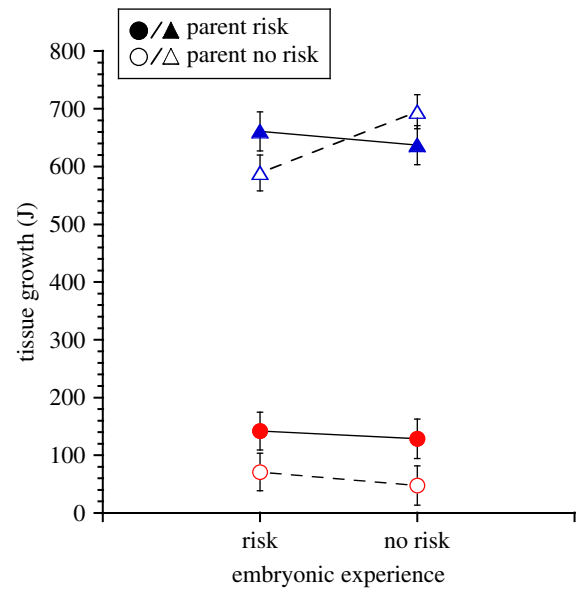


Figure 2. Mean (\pm s.e.) tissue growth (Joules, J) of offspring *Nucella lapillus* in the presence (circles) and absence (triangles) of current predation risk from the green crab *Carcinus maenas*. Offspring experienced the presence and absence of green crab predation risk as embryos and were produced by parents that experienced the presence (filled symbols) and absence (open symbols) of green crab predation risk. $n = 8$ for all treatment combinations except: – parental risk/– embryonic risk/+ current risk and + parental risk/+ embryonic risk/+ current risk. (Online version in colour.)

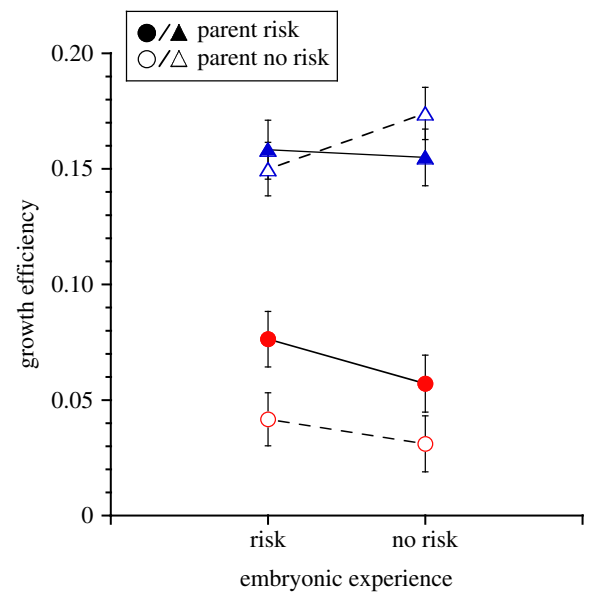


Figure 3. Mean (\pm s.e.) growth efficiency of offspring *Nucella lapillus* in the presence (circles) and absence (triangles) of current predation risk from the green crab *Carcinus maenas*. Offspring experienced the presence and absence of green crab predation risk as embryos and were produced by parents that experienced the presence (filled symbols) and absence (open symbols) of green crab predation risk. The three-way interaction is not significant here (parental \times embryonic \times current: $p = 0.5$), but is shown for ease of comparison to figure 2. $n = 8$ for all treatment combinations except: – parental risk/– embryonic risk/+ current risk and + parental risk/+ embryonic risk/+ current risk. (Online version in colour.)

embryonic risk experience on offspring tissue growth in the presence of current risk (LS contrast: $p = 0.5$).

Offspring also grew less efficiently in the presence of current risk ($F_{1,53.2} = 312.9$, $p < 0.0001$; figure 3), but offspring of

risk-experienced parents were 88% more efficient than the offspring of risk-naïve parents (parental experience \times current risk: $F_{1,53.2} = 8.8$, $p = 0.004$, LS contrast: $p = 0.03$; figure 3) in the presence of current risk. There was no effect of parental risk experience on growth efficiency in the absence of current risk (LS contrast: $p = 0.7$). There was also an interaction between embryonic risk experience and current risk ($F_{1,53.0} = 4.2$, $p = 0.046$), but post hoc tests revealed no differences between means (LS contrasts, presence of embryonic risk: $p = 0.2$, absence of embryonic risk: $p = 0.3$). Growth efficiency was not affected by embryonic risk experience ($F_{1,1.3} = 0.05$, $p = 0.8$). Finally, none of the random effects, including parent pair, affected offspring growth or growth efficiency (electronic supplementary material, table S3).

4. Discussion

In the presence of current green crab predation risk, *Nucella* offspring used refuges more and had lower foraging activity, growth and growth efficiency than those in the absence of current risk (figures 1–3). These results are consistent with previous work where *Nucella* increase their antipredator behaviour and exhibit reduced performance in the presence of risk (e.g. [45]). Parental experience with predation risk, however, largely reversed these effects—offspring of risk-experienced parents spent less time in refuge, grew more and had higher growth efficiency in the presence of current risk than offspring of risk-naïve parents. These results support our previous work that parental experience with green crab predation risk improves the performance of 1-year-old *Nucella* offspring [47]. Importantly, while our risk manipulation treatment exposed offspring to constant predation risk throughout the experiment, *Nucella* that are intermittently (e.g. 25% of the time) exposed to risk from *Carcinus* experience similar fitness consequences as those in constant risk [59].

In the presence of current risk, offspring of risk-experienced parents spent less time in refuge habitats than offspring of risk-naïve parents (figure 1), suggesting that parental effects impact how offspring manage risk. Parental risk experience may increase the reliability of information communicated by predator cues to offspring and thus enhance the ability of offspring to validate the potential persistence of predation risk in their environment. Temporal variation in risk can be a strong driver of prey behaviour; both theoretical and empirical work suggest that prey which are consistently exposed to high risk display weaker antipredator behaviour than those exposed to shorter, more variable risk (risk allocation hypothesis [53,60]). Parental experience with risk may also influence how offspring perceive variability in their risk environment by decreasing uncertainty about future risk conditions. Hence, offspring may be emboldened to leave a refuge even after short risk exposure because they judge the costs of prolonged antipredator behaviour (e.g. reduced foraging) to be higher than its benefits. By contrast, the offspring of risk-naïve parents may have greater uncertainty about present and future risk conditions and, therefore, remain in refuge longer even as the costs of doing so escalate. Prey will only confront such decisions when the costs of antipredator behaviour are sufficiently great, which is predicted to occur only after prolonged periods of risk exposure such as those in our experiment (25 days [33,61]). Therefore, while our results differ from studies in other

systems showing that offspring of risk-experienced parents display greater antipredator behaviour during very brief risk exposures [62,63], the costs of such behaviours probably manifest only after longer periods of risk exposure.

Interestingly, parental experience with risk did not impact offspring foraging activity in either the presence or absence of current risk (electronic supplementary material, figure S1) despite differences in refuge use. This similarity in foraging may have emerged because the offspring of risk-experienced and risk-naïve parents employed different foraging strategies. For example, offspring of risk-naïve parents may move back and forth between the risky and refuge habitats, whereas offspring of risk-experienced parents remain in the risky habitat. However, more frequent behavioural observations would be necessary to explore this hypothesis. While other experiments in this system have detected a clear trade-off between foraging and hiding (e.g. [44,64]), the current study indicates that this issue can be quite nuanced. Despite this observed similarity in foraging, the offspring of risk-experienced parents grew more tissue than offspring of risk-naïve parents (figure 2) because they had higher growth efficiency in the presence of current risk (figure 3). These results suggest that parental effects operated through physiological changes in offspring in response to current predation risk. Exposure to predation risk often has negative effects on prey growth efficiency [46,65] because prey allocate energy away from growth to support increased respiration and other physiological pathways that mitigate the impacts of stress [66,67]. Elsewhere [42] we have shown that parental risk experience probably contributes to reduced respiration rates in *Nucella* embryos that are exposed to green crab risk cues, suggesting that offspring of risk-experienced parents exhibit a weaker stress response or require less energy to do so.

Parental effects often operate through epigenetic modifications in offspring that make genes more or less accessible to transcription [68,69], which may impact the energy required for gene expression. We do not know if parental effects act through epigenetic modifications in this system and this intriguing hypothesis awaits future work. Moreover, despite apparent improvements in offspring growth and physiology, we have yet to establish whether these changes reflect adaptive parental effects because we have not observed their impacts on the reproductive output of parents or the survival of offspring. The reduction in antipredator behaviour among offspring of risk-experienced parents could improve offspring survival under extended periods of risk by reducing the risk of starvation (see above), which may outweigh the risk of being eaten. In any case, the positive relationship between individual size and fecundity observed in many systems (e.g. [70]) suggests that parental experience with risk may increase the reproductive output of subadult offspring *Nucella* in the presence of current risk.

Importantly, parental effects only operated when offspring were exposed to current risk, and such context-dependency is common in the expression of parental effects [17,31,71,72]. Parental effects may be more beneficial to offspring when parental and offspring environments are similar [73–75] and when offspring face adverse conditions [76] such as those under predation risk. Our results did not reveal costs of parental risk experience for offspring in the absence of current risk, which is surprising given previous work showing that parental or early life effects can be maladaptive when they do not accurately predict future offspring

environmental conditions [73–75]. However, the adaptive value of parental effects may depend upon the specific mechanism through which they impact offspring fitness. If parental effects act to reduce the physiological costs of predation risk in their offspring as suggested by our results, such changes would be unnecessary in benign, risk-free conditions. It is also possible that despite our monitoring of offspring traits across multiple stages of life history (this study and [42]), we have yet to isolate the trait or stage of ontogeny where such costs are evident. Finally, the costs of parental effects in this system may only appear in the parent generation [19], which was not monitored here.

Although embryonic experience with predation risk did not significantly affect offspring refuge use, foraging activity or growth efficiency in either the presence or absence of current risk, it did affect offspring tissue growth in the absence of current risk—offspring that were exposed to predation risk as embryos produced less tissue than offspring that were not exposed to risk as embryos (figure 2). These reductions in offspring tissue growth in response to embryonic risk experience were not driven by changes in growth efficiency as they were for offspring based on parental risk experience (figure 3). Interestingly, however, these size patterns correspond with those that we found for offspring at emergence in our earlier work—offspring of risk-naïve parents emerged smaller from development if they were exposed to risk as embryos [42]. We hypothesize that a silver spoon effect [77,78] may be operating: offspring that experienced relatively benign, risk-free conditions as embryos were relatively better off and, therefore, grew more as 1-year-old adults than those that experienced stressful, risky conditions as embryos. Hence, low-stress conditions early in life may have lasting and positive effects on offspring performance later in life. It is possible that these changes in offspring growth based on embryonic risk experiences impact the willingness of offspring to forage. While only correlative, our results suggest that in the absence of current risk, offspring that experienced risk as embryos consumed on average approximately 125 fewer Joules *per capita* and produced on average approximately 106 J less tissue than offspring that did not experience risk as embryos. Foraging can be inherently risky regardless of habitat [79]; for example, consuming a mussel can leave *Nucella* vulnerable for extended periods of time [80]. Offspring that experienced risk as embryos may be more averse to such risk, regardless of their current risk conditions, which may have driven the slight reductions in tissue growth.

We found that embryonic risk experience had no effect on *Nucella* offspring performance in the presence of current risk, possibly because the impact of embryonic risk experience on offspring growth was relatively small compared to the effect of current risk exposure. Indeed, the overall magnitude of the embryonic effect was much smaller than the effects of either parental or current risk exposure: when operating, parental and current risk experiences suppressed offspring growth by 100% and 85%, respectively, whereas embryonic experience with risk-reduced growth by only 17%. Hence, even if embryonic effects were operating on offspring fitness in the presence of current risk, they may not have been substantial enough for us to detect.

The parents of the offspring used in this experiment were collected directly from the field, which may introduce heterogeneity in parental risk experiences. We attempted to minimize these potential effects by collecting parents from a

relatively small spatial area (see Material and methods) and including ‘family’ in our statistical models. However, we recognize that all organisms are probably influenced by past experiences, either directly or indirectly (e.g. parental or grand-parental) and that the response of parents to risk may further depend on experiences not manipulated in our experiment. Nevertheless, our results show a strong effect of parental experience despite these potential differences, suggesting that parental experience with risk prior to mating can have important impacts on offspring performance.

Our results reveal that the effects of parental and embryonic experiences are largely independent later in ontogeny. By contrast, our previous work [42] (on siblings of the offspring studied here) found that these effects early in offspring ontogeny operated synergistically to impact offspring size at emergence. Embryonic effects may be more influential during early life history because embryonic cues are deemed more reliable during early stages of ontogeny. Theory predicts that the use of environmental information should decline over time as organisms obtain more recent, and therefore relevant, information [27]. One might expect this rationale to apply to parental effects, but our results suggest that parental effects continued to have strong impacts on offspring throughout ontogeny. Hence, it appears that offspring may ‘trust’ their parents more than themselves when evaluating risk. Parental experience is the earliest possible source of information for offspring and, therefore, may inherently have more profound effects on phenotypic outcomes [12,27,81]. Furthermore, the strength and persistence of parental effects may be driven by their influence on offspring physiology, which may have more significant consequences for offspring than behavioural changes alone.

In summary, our results suggest that parental and embryonic experiences can affect offspring performance, but that their relative importance and tendency to interact may change over time based on environmental context. The persistence of parental and embryonic effects may also depend on the mechanisms through which they operate, so an exploration of the pathways that drive such effects is probably important when examining the role of parental and embryonic effects across an individual’s lifetime. We suggest that changes in physiological performance may be more influential on offspring lifetime fitness than changes in foraging rates. Our results demonstrate that attention to the complexity of offspring responses based on parental and embryonic experiences will be essential to robustly predict how natural populations and communities will respond to changing environments.

Ethics. This work was conducted in accordance with the guidelines of the Association for the Study of Animal Behavior and the animal care guidelines of Northeastern University’s Institutional Animal Care and Use Committee (IACUC).

Data accessibility. Data are available at the Dryad Digital Repository: <http://dx.doi.org/10.5061/dryad.ks622> [58].

Authors’ contributions. Both authors conceived of the study, S.C.D. designed and coordinated the study, carried out the statistical analyses and drafted the manuscript. Both authors contributed equally to manuscript revisions and give final approval for publication.

Competing interests. We have no competing interests.

Funding. This study was generously supported by National Science Foundation grants to G.C.T. (IOS-1557901, OCE-0963010, Academic Research Infrastructure Recovery and Reinvestment Program, and OCE-1458150).

Acknowledgements. We thank Erin Bucci, Rachel Dowley, Erin Sayre and Sydney Stenquist for their enormous help with all aspects of

experimental set-up and maintenance, Kyle Pepperman and the Downeast Institute for supplying juvenile mussels, Jessica Torossian for mussel transport, Eric Sanford for experimental advice, Catherine

Matassa for intellectual feedback and three anonymous reviewers for their helpful comments. This is part of the PhD dissertation of S.C.D. It is contribution no. 364 from the Marine Science Center.

References

- West-Eberhard MJ. 1989 Phenotypic plasticity and the origins of diversity. *Ann. Rev. Ecol. Syst.* **20**, 249–278. (doi:10.1146/annurev.es.20.110189.001341)
- Pigliucci M. 2001 *Phenotypic plasticity: beyond nature and nurture*. Baltimore, MD: Johns Hopkins University Press.
- Via S, Gomulkiewicz R, De Jong G, Scheiner SM, Schlichting CD, Van Tienderen PH. 1995 Adaptive phenotypic plasticity: consensus and controversy. *Trends Ecol. Evol.* **10**, 212–217. (doi:10.1016/S0169-5347(00)89061-8)
- Schmitz OJ. 2003 Top predator control of plant biodiversity and productivity in an old field ecosystem. *Ecol. Lett.* **6**, 156–163. (doi:10.1046/j.1461-0248.2003.00412.x)
- Miner BG, Sultan SE, Morgan SG, Padilla DK, Relyea RA. 2005 Ecological consequences of phenotypic plasticity. *Trends Ecol. Evol.* **20**, 685–692. (doi:10.1016/j.tree.2005.08.002)
- O'Connor CM, Norris DR, Crossin GT, Cooke SJ. 2014 Biological carryover effects: linking common concepts and mechanisms in ecology and evolution. *Ecosphere* **5**, 1–11. (doi:10.1890/ES13-00388.1)
- West-Eberhard MJ. 2003 *Developmental plasticity and evolution*. Oxford, UK: Oxford University Press.
- Williams TD. 1994 Intraspecific variation in egg size and egg composition in birds: effects on offspring fitness. *Biol. Rev.* **69**, 35–59. (doi:10.1111/j.1469-185X.1994.tb01485.x)
- Warkentin KM. 1995 Adaptive plasticity in hatching age: a response to predation risk trade-offs. *Proc. Natl Acad. Sci. USA* **92**, 3507–3510. (doi:10.1073/pnas.92.8.3507)
- Metcalfe NB, Monaghan P. 2001 Compensation for a bad start: grow now, pay later? *Trends Ecol. Evol.* **16**, 254–260. (doi:10.1016/S0169-5347(01)02124-3)
- De Block M, Stoks R. 2005 Fitness effects from egg to reproduction: bridging the life history transition. *Ecology* **86**, 185–197. (doi:10.1890/04-0116)
- English S, Fawcett TW, Higginson AD, Trimmer PC, Uller T, Gaillard J-M, Bronstein JL. 2016 Adaptive use of information during growth can explain long-term effects of early life experiences. *Am. Nat.* **187**, 620–632. (doi:10.1086/685644)
- Pechenik JA. 2006 Larval experience and latent effects: metamorphosis is not a new beginning. *Integr. Comp. Biol.* **46**, 323–333. (doi:10.1093/icb/ijc028)
- Naguib M, Gil D. 2005 Transgenerational body size effects caused by early developmental stress in zebra finches. *Biol. Lett.* **1**, 95–97. (doi:10.1098/rsbl.2004.0277)
- Bernardo J. 1996 Maternal effects in animal ecology. *Am. Zool.* **36**, 83–105. (doi:10.1093/icb/36.2.83)
- Mousseau TA, Fox CW. 1998 *Maternal effects as adaptations*. Oxford, UK: Oxford University Press.
- Räsänen K, Kruuk L. 2007 Maternal effects and evolution at ecological time-scales. *Funct. Ecol.* **21**, 408–421. (doi:10.1111/j.1365-2435.2007.01246.x)
- Uller T. 2008 Developmental plasticity and the evolution of parental effects. *Trends Ecol. Evol.* **23**, 432–438. (doi:10.1016/j.tree.2008.04.005)
- Marshall DJ, Uller T. 2007 When is a maternal effect adaptive? *Oikos* **116**, 1957–1963. (doi:10.1111/j.2007.0030-1299.16203.x)
- Burgess SC, Marshall DJ. 2014 Adaptive parental effects: the importance of estimating environmental predictability and offspring fitness appropriately. *Oikos* **123**, 769–776. (doi:10.1111/oik.01235)
- Agrawal AA, Laforsch C, Tollrian R. 1999 Transgenerational induction of defences in animals and plants. *Nature* **401**, 60–63. (doi:10.1038/43425)
- Fox CW, Thakar MS, Mousseau TA. 1997 Egg size plasticity in a seed beetle: an adaptive maternal effect. *Am. Nat.* **149**, 149–163. (doi:10.1086/285983)
- Miller GM, Watson S-A, Donelson JM, McCormick MI, Munday PL. 2012 Parental environment mediates impacts of increased carbon dioxide on a coral reef fish. *Nat. Clim. Change* **2**, 858–861. (doi:10.1038/nclimate1599)
- LaMontagne JM, McCauley E. 2001 Maternal effects in *Daphnia*: what mothers are telling their offspring and do they listen? *Ecol. Lett.* **4**, 64–71. (doi:10.1046/j.1461-0248.2001.00197.x)
- Naguib M, Nemitz A, Gil D. 2006 Maternal developmental stress reduces reproductive success of female offspring in zebra finches. *Proc. R. Soc. B* **273**, 1901–1905. (doi:10.1098/rspb.2006.3526)
- Herman JJ, Sultan SE, Horgan-Kobelski T, Riggs C. 2012 Adaptive transgenerational plasticity in an annual plant: grandparental and parental drought stress enhance performance of seedlings in dry soil. *Integr. Comp. Biol.* **52**, 77–88. (doi:10.1093/icb/ics04)
- Dufty AM, Clobert J, Møller AP. 2002 Hormones, developmental plasticity and adaptation. *Trends Ecol. Evol.* **17**, 190–196. (doi:10.1016/S0169-5347(02)02498-9)
- Lindholm AK, Hunt J, Brooks R. 2006 Where do all the maternal effects go? Variation in offspring body size through ontogeny in the live-bearing fish *Poecilia parae*. *Biol. Lett.* **2**, 586–589. (doi:10.1098/rsbl.2006.0546)
- Andree SR, Feiner ZS, Bledsoe JW, Cragun AM, Höök TO. 2015 Ontogenetic variability of maternal effects in an iteroparous fish. *Ecol. Freshw. Fish* **24**, 384–396. (doi:10.1111/eff.12153)
- Pahkala M, Laurila A, Merilä J. 2001 Carry-over effects of ultraviolet-B radiation on larval fitness in *Rana temporaria*. *Proc. R. Soc. B* **268**, 1699–1706. (doi:10.1098/rspb.2001.1725)
- Marshall DJ. 2008 Transgenerational plasticity in the sea: context-dependent maternal effects across the life history. *Ecology* **89**, 418–427. (doi:10.1890/07-0449.1)
- Sih A, Petranka JW, Kats LB. 1988 The dynamics of prey refuge use: a model and tests with sunfish and salamander larvae. *Am. Nat.* **132**, 463–483. (doi:10.1086/284865)
- Lima SL, Dill LM. 1990 Behavioral decisions made under the risk of predation: a review and prospectus. *Can. J. Zool.* **68**, 619–640. (doi:10.1139/z90-092)
- Kats LB, Dill LM. 1998 The scent of death: chemosensory assessment of predation risk by prey animals. *Ecoscience* **5**, 361–394. (doi:10.1080/11956860.1998.11682468)
- McCauley SJ, Rowe L, Fortin M-J. 2011 The deadly effects of 'nonlethal' predators. *Ecology* **92**, 2043–2048. (doi:10.1890/11-0455.1)
- Sheriff MJ, Krebs CJ, Boonstra R. 2010 The ghosts of predators past: population cycles and the role of maternal programming under fluctuating predation risk. *Ecology* **91**, 2983–2994. (doi:10.1890/09-1108.1)
- Matassa CM, Trussell GC. 2011 Landscape of fear influences the relative importance of consumptive and nonconsumptive predator effects. *Ecology* **92**, 2258–2266. (doi:10.1890/11-0424.1)
- Schmitz OJ, Hawlena D, Trussell GC. 2010 Predator control of ecosystem nutrient dynamics. *Ecol. Lett.* **13**, 1199–1209. (doi:10.1111/j.1461-0248.2010.01511.x)
- Coslovsky M, Richner H. 2011 Predation risk affects offspring growth via maternal effects. *Funct. Ecol.* **25**, 878–888. (doi:10.1111/j.1365-2435.2011.01834.x)
- Peckarsky BL, Taylor BW, McIntosh AR, McPeck MA, Lytle DA. 2001 Variation in mayfly size at metamorphosis as a developmental response to risk of predation. *Ecology* **82**, 740–757. (doi:10.1890/0012-9658(2001)082[0740:VIMSAM]2.0.CO;2)
- McGhee KE, Pintor LM, Suhr EL, Bell AM. 2012 Maternal exposure to predation risk decreases offspring antipredator behaviour and survival in threespined stickleback. *Funct. Ecol.* **26**, 932–940. (doi:10.1111/j.1365-2435.2012.02008.x)
- Donelan SC, Trussell GC. 2018 Synergistic effects of parental and embryonic exposure to predation risk on prey offspring size at emergence. *Ecology* **99**, 68–78. (doi:10.1002/ecy.2067)
- Trussell GC, Ewanchuk PJ, Matassa CM. 2006 Habitat effects on the relative importance of trait- and density-mediated indirect interactions. *Ecol. Lett.* **9**, 1245–1252. (doi:10.1111/j.1461-0248.2006.00981.x)

44. Donelan SC, Grabowski JH, Trussell GC. 2017 Refuge quality impacts the strength of nonconsumptive effects on prey. *Ecology* **98**, 403–411. (doi:10.1002/ecy.1647)
45. Matassa CM, Donelan SC, Luttbeg B, Trussell GC. 2016 Resource levels and prey state influence antipredator behavior and the strength of nonconsumptive predator effects. *Oikos* **125**, 1478–1488. (doi:10.1111/oik.03165)
46. Trussell GC, Ewanchuk PJ, Matassa CM. 2006 The fear of being eaten reduces energy transfer in a simple food chain. *Ecology* **87**, 2979–2984. (doi:10.1890/0012-9658(2006)87[2979:TFOBER]2.0.CO;2)
47. Donelan SC, Trussell GC. 2015 Parental effects enhance risk tolerance and performance in offspring. *Ecology* **96**, 2049–2055. (doi:10.1890/14-1773.1)
48. Etter RJ. 1989 Life history variation in the intertidal snail *Nucella lapillus* across a wave-exposure gradient. *Ecology* **70**, 1857–1876. (doi:10.2307/1938118)
49. Crothers J. 1985 Dogwhelks: an introduction to the biology of *Nucella lapillus* (L.). *Field Stud.* **6**, 291–360.
50. Hughes RN. 1972 Annual production of two Nova Scotian populations of *Nucella lapillus* (L.). *Oecologia* **8**, 356–370. (doi:10.1007/BF00367538)
51. Large SI, Smee DL. 2010 Type and nature of cues used by *Nucella lapillus* to evaluate predation risk. *J. Exp. Mar. Biol. Ecol.* **396**, 10–17. (doi:10.1016/j.jembe.2010.10.005)
52. Palmer AR. 1982 Growth in marine gastropods: a non-destructive technique for independently measuring shell and body weight. *Malacologia* **23**, 63–74.
53. Matassa CM, Trussell GC. 2014 Prey state shapes the effects of temporal variation in predation risk. *Proc. R. Soc. B* **281**, 20141952. (doi:10.1098/rspb.2014.1952)
54. Burrows MT, Hughes RN. 1990 Variation in growth and consumption among individuals and populations of dogwhelks, *Nucella lapillus*: a link between foraging behaviour and fitness. *J. Anim. Ecol.* **59**, 723–742. (doi:10.2307/4891)
55. Elner RW, Hughes RN. 1978 Energy maximization in the diet of the shore crab, *Carcinus maenas*. *J. Anim. Ecol.* **47**, 103–116. (doi:10.2307/3925)
56. Palmer AR. 1992 Calcification in marine molluscs: how costly is it? *Proc. Natl Acad. Sci. USA* **89**, 1379–1382. (doi:10.1073/pnas.89.4.1379)
57. Zuur AF, Ieno EN, Walker NJ, Saveliev AA, Smith GM. 2009 *Mixed effects models and extensions in ecology with R*. New York, NY: Springer.
58. Donelan SC, Trussell GC. 2018 Data from: Parental and embryonic experiences with predation risk affect prey offspring behavior and performance. Dryad Digital Repository. (<http://dx.doi.org/10.5061/dryad.ks622>)
59. Trussell GC, Matassa CM, Luttbeg B. 2011 The effects of variable predation risk on foraging and growth: less risk is not necessarily better. *Ecology* **92**, 1799–1806. (doi:10.1890/10-2222.1)
60. Lima SL, Bedneko PA. 1999 Temporal variation in danger drives antipredator behavior: the predation risk allocation hypothesis. *Am. Nat.* **153**, 649–659. (doi:10.1086/303202)
61. Werner EE, Gilliam JF, Hall DJ, Mittelbach GG. 1983 An experimental test of the effects of predation risk on habitat use in fish. *Ecology* **64**, 1540–1548. (doi:10.2307/1937508)
62. Giesing ER, Suski CD, Warner RE, Bell AM. 2010 Female sticklebacks transfer information via eggs: effects of maternal experience with predators on offspring. *Proc. R. Soc. B* **278**, 1753–1759. (doi:10.1098/rspb.2010.1819)
63. Storm JJ, Lima Steven L. 2010 Mothers forewarn offspring about predators: a transgenerational maternal effect on behavior. *Am. Nat.* **175**, 382–390. (doi:10.1086/650443)
64. Trussell GC, Ewanchuk PJ, Matassa CM. 2008 Resource identity modifies the influence of predation risk on ecosystem function. *Ecology* **89**, 2798–2807. (doi:10.1890/08-0250.1)
65. McPeck MA, Grace M., Richardson JM. 2001 Physiological and behavioral responses to predators shape the growth/predation risk trade-off in damselfishes. *Ecology* **82**, 1535–1545. (doi:10.1890/0012-9658(2001)082[1535:PABRTP]2.0.CO;2)
66. Slos S, Stoks R. 2008 Predation risk induces stress proteins and reduces antioxidant defense. *Funct. Ecol.* **22**, 637–642. (doi:10.1111/j.1365-2435.2008.01424.x)
67. Hawlena D, Schmitz OJ. 2010 Physiological stress as a fundamental mechanism linking predation to ecosystem functioning. *Am. Nat.* **176**, 537–556. (doi:10.1086/656495)
68. Jablonka E, Lamb MJ. 1998 Epigenetic inheritance in evolution. *J. Evol. Biol.* **11**, 159–183. (doi:10.1046/j.1420-9101.1998.11020159.x)
69. Holeski LM, Jander G, Agrawal AA. 2012 Transgenerational defense induction and epigenetic inheritance in plants. *Trends Ecol. Evol.* **27**, 618–626. (doi:10.1016/j.tree.2012.07.011)
70. Honěk A. 1993 Intraspecific variation in body size and fecundity in insects: a general relationship. *Oikos* **66**, 483–492. (doi:10.2307/3544943)
71. Badyaev AV, Uller T. 2009 Parental effects in ecology and evolution: mechanisms, processes and implications. *Phil. Trans. R. Soc. B* **364**, 1169–1177. (doi:10.1098/rstb.2008.0302)
72. Plaistow S, Benton T. 2009 The influence of context-dependent maternal effects on population dynamics: an experimental test. *Phil. Trans. R. Soc. B* **364**, 1049–1058. (doi:10.1098/rstb.2008.0251)
73. Love OP, Williams TD. 2008 The adaptive value of stress-induced phenotypes: effects of maternally derived corticosterone on sex-biased investment, cost of reproduction, and maternal fitness. *Am. Nat.* **172**, E135–E149. (doi:10.1086/590959)
74. Monaghan P. 2008 Early growth conditions, phenotypic development and environmental change. *Phil. Trans. R. Soc. B* **363**, 1635–1645. (doi:10.1098/rstb.2007.0011)
75. Sheriff M, Love O. 2013 Determining the adaptive potential of maternal stress. *Ecol. Lett.* **16**, 271–280. (doi:10.1111/ele.12042)
76. Räsänen K, Laurila A, Merilä J. 2005 Maternal investment in egg size: environment and population-specific effects on offspring performance. *Oecologia* **142**, 546–553. (doi:10.1007/s00442-004-1762-5)
77. Madsen T, Shine R. 2000 Silver spoons and snake body sizes: prey availability early in life influences long-term growth rates of free-ranging pythons. *J. Anim. Ecol.* **69**, 952–958. (doi:10.1111/j.1365-2656.2000.00477.x)
78. Grafen A. 1988 On the uses of data on lifetime reproductive success. In *Reproductive success studies of individual variation in contrasting breeding systems* (ed. TH Clutton-Brock), pp. 454–471. Chicago, IL: University of Chicago.
79. Brown JS. 1999 Vigilance, patch use and habitat selection: foraging under predation risk. *Evol. Ecol. Res.* **1**, 49–71. (doi:10.1.1.489.6835)
80. Miller LP. 2013 The effect of water temperature on drilling and ingestion rates of the dogwhelk *Nucella lapillus* feeding on *Mytilus edulis* mussels in the laboratory. *Mar. Biol.* **160**, 1489–1496. (doi:10.1007/s00227-013-2202-z)
81. Fawcett TW, Frankenhuis WE. 2015 Adaptive explanations for sensitive windows in development. *Front. Zool.* **12**, S3. (doi:10.1186/1742-9994-12-S1-53)